## Stock Assessment of red cod (Salilota australis) in the Falkland Islands



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## Summary

The red cod Total Allowable Catch (TAC) for 2023 is set at 1,439 tonnes ( t ). Following recommendations of the MacAlister Elliott \& Partners external review, this TAC was calculated according to the ICES category 3-2/3 rule, in which next year's advised TAC is proportioned by the mean biomass estimate of the two most recent years divided by the mean biomass estimate of the three previous years.

Red cod commercial catches in Falkland Islands licenced fisheries were 1,187 t in 2021; below the average catch over the past 10 years.

Red cod commercial CPUE in the Falklands Interim Conservation Zone varied between 1990 and 2021, with the CPUE in 2020 and 2021 being the lowest in the time series.

Length-based indicators suggest that conservation of immature fish, large individuals, mega-spawners, optimal yield, and MSY have been poor for over a decade. Length and age analyses showed no significant change of asymptotic length from 2002 through 2019 and of length at 50\% maturity of females from 1992 through 2021. Comparison of length at 50\% maturity and catch at length revealed that red cod were mainly caught before reaching maturity, which can reduce stock sustainability. Poor indicators for large individuals and for mega-spawners, and the higher proportion of immature individuals within the catch over the past decade may be in part because of fishing vessels being prohibited from targeting spawning individuals since 2010.

## Introduction

Red cod (Salilota australis, Moridae) is a demersal fish that inhabits shelf and slope subtropical, temperate and sub-Antarctic waters, at $30-1,000 \mathrm{~m}$ depth, of the Southeast Pacific (from $33^{\circ} \mathrm{S}$ to the southern tip of South America) and Southwest Atlantic ( $40^{\circ} \mathrm{S}$ to $55^{\circ} \mathrm{S}$ ) in the Patagonian region of Chile and Argentina, and in Falkland Islands and Strait of Magellan waters (Cohen et al. 1990; Arkhipkin et al. 2010, 2012). However, no studies have addressed connectivity of this species between the Southeast Pacific and the Southwest Atlantic. Red cod carries out winter (July through September) migrations on the Patagonian Shelf in the Southwest Atlantic (Arkhipkin et al. 2010). Spawning aggregations occur at 180-200 m depths in the periphery of two upwelling areas created by the cold-water Falkland Current to the south and southwest of West Falkland between August and October (Arkhipkin et al. 2010; Brickle et al. 2011). After spawning, red cod migrate back to their feeding grounds (Arkhipkin et al. 2010). Based on the migratory behaviour of red cod in the region it is assumed that Falklands and Argentine fisheries catch the same stock. This species is considered a valuable by-catch by these two nations and most of the catch is historically taken by the Falkland Islands fishery (Falkland Islands Government ${ }^{\text {a }}$; Falkland Islands Government 2021), compared with the Argentina fishery (Argentine Government ${ }^{\text {b }}$; Sánchez et al. 2012; Navarro et al. 2014, 2019).

## Methods

## ICES advice rules

In 2020, red cod was included in a Falkland Islands Government finfish stock assessment and management review conducted by MacAlister Elliott \& Partners Ltd, UK (MEP 2020). The MEP report recommended stock assessments for most commercial finfish species to be based on the ICES advice rules (ICES 2012, 2018a), referencing applicable categories of data availability and quality; for red cod, the advice was to calculate Total Allowable Catch (TAC) using the ICES category 3, as a species for which commercial landings data and abundance indices from surveys or the fishery are available. MEP (2020) also recommended exploring ancillary stock status information from ICES data limited methods such as lengthbased indicators. A Length-Based Indicator method (LBI) has been used since 2021 by the

[^0]Falkland Islands Fisheries Department (FIFD) to provide a suite of indicators for several commercial finfish species based on combinations of catch-at-size distributions, and lifehistory parameters such as Linf (asymptotic length; Haddon 2001) and L50 (length at 50\% maturity; Cope \& Punt 2009). Otolith growth increments of Falkland Islands red cod have been read at the National Marine Fisheries Research Institute (MFRI) in Gdynia, Poland. Red cod age data in the FIFD database have been rated 'with caution' (Lee et al. 2020) but appear reliable; therefore, LBI was implemented.

## ICES Category 3 Total Allowable Catch

For category 3 the common assessment method uses a $2 / 3$ rule, in which next year's advised TAC is calculated taking into account the most recent five years of the index. By this rule, a ratio of the mean of the last two years over the mean of the first three years' index of the five-year series is multiplied against the current year TAC (Ramos \& Winter 2021) to generate next year's TAC (MEP 2020). However, each year's survey biomass estimate is subject to more or less uncertainty depending on the distribution of catches (Ramos \& Winter 2022). To reflect uncertainty, biomass estimates were weighted by their inverse variance (Marín-Martínez \& Sánchez-Meca 2010). Variances of red cod biomass were estimated from each year's 10,000 randomized re-samples (2015 $=844641348.4 ; 2016=1008476074.0 ; 2017$ $=322677148.5 ; 2018=823259842.4 ; 2019=485141808.8 ; 2020=34324611.8 ; 2021=$ 54115516.4; $2022=4401033099.0$ ). TAC is further limited to an 'uncertainty cap' of $\pm 20 \%$ (ICES 2018a) with respect of the TAC set for the current year ( $\mathrm{TAC}_{2022}=1,199 \mathrm{t}$; Ramos \& Winter 2021). TAC was calculated as follows:

$$
\left.T A C_{-} 3_{2023}=T A C_{-} 3_{2022} \times \frac{\left(\sum_{y=2021}^{2022} B_{y} \times 1 / V a r_{y}\right) /\left(\sum_{y=2021}^{2022} 1 / V a r_{y}\right)}{\left(\sum_{y=2018}^{2020} B_{y} \times 1 / V a r_{y}\right) /\left(\sum_{y=2018}^{2020} 1 / V a r_{y}\right)} \right\rvert\, \pm 20 \%
$$

Where $\mathrm{B}=$ February surveys biomass ( t ), and Var = Variance of February surveys biomass.

## Commercial catch and CPUE

Commercial fishing around the Falkland Islands was not distinguished from other parts of the Southwest Atlantic prior to 1982 and catch data by species were recorded systematically from 1987 only (Falkland Islands Government 1989). Therefore, total red cod catch data were examined from 1987 to 2021 from the Falkland Islands (Falkland Islands Government ${ }^{\text {a }}$; Falkland Islands Government 2021), and Argentina (Argentine Government ${ }^{\text {b }}$; Sánchez et al. 2012; Navarro et al. 2014, 2019). LOESS (span $=0.75$, degree $=2$ ) was implemented to examine the pattern of the association between Falkland Islands and Argentina commercial annual catches of red cod from 1987 through 2021. Commercial catches and discard of red cod were examined by licence type for 2021 in the FICZ.

CPUE was calculated as the sum of red cod catches divided by the sum of effort (trawl hours); annual CPUE, monthly CPUE through the time series, and the monthly distribution of the CPUE in the FICZ during 2021 were examined. Annual CPUE was calculated from bottom trawl finfish (A-, G-, and W-licences) vessels with fishing activity through the FICZ from 1990 through 2021. Monthly CPUE was calculated from finfish vessels with fishing activity through the FICZ from 1990 through 2020, and for 2021. CPUE was calculated from $A-, G-$, and $W-$ licences because these contribute most of the red cod catches throughout the years. LOESS (span $=0.75$, degree $=2$ ) was implemented to examine the patterns of annual and monthly CPUE.

## Survey biomass estimates

Biomass estimates and the spatial distribution of red cod were examined from joint surveys (groundfish and Patagonian squid Doryteuthis gahi pre-season surveys) carried out in February 2010, 2011, and 2015 - 2022 in Falkland Islands waters (Ramos \& Winter 2022). Biomass ratios between the most recent February surveys (2022) and the first February surveys (2010) were estimated as a proxy of the change in biomass over time. Significance of difference and $95 \%$ confidence intervals of the change in biomass were computed from the randomized re-samples of the survey biomass estimates (Ramos \& Winter 2022). A trend of the biomass time series from 2010 to 2022 was calculated using LOESS (span = 1, degree $=2$ ).

[^1]Biomass estimates, the spatial distribution of red cod, and biomass ratios were also examined following Ramos \& Winter (2022) from joint surveys (groundfish and Patagonian squid pre-season surveys) carried out during July 2017 (Gras et al. 2017; Winter et al. 2017) and July 2020 (Randhawa et al. 2020; Winter et al. 2020). The July surveys were conducted for the primary purpose of assessing common hake (Gras et al. 2017; Randhawa et al. 2020), and are presented as an additional comparative proxy for abundance patterns, with the caveat that these would likely reflect variability in the migratory timing of red cod.

## Length and age analyses

## Length Based Indicators

ICES $(2015,2018 b)$ recommends the LBI method which provides a suite of indicators based on combinations of catch-at-size distributions, life-history parameters such as Linf (asymptotic length; Haddon 2001) and L50 (length at 50\% maturity; Cope \& Punt 2009). Linf and L50 parameters were assessed for females and males separately as red cod have sexually dimorphic growth (Brickle et al. 2011; Chong Follert et al. 2017).

LBI method was applied to all years from which red cod length and age data were available and reported as random samples (FIFD database codes R and S), i.e., years 1992 to 2021 for length data, and years 2002 to 2019 for age data. Because finfish trawls are restricted to larger meshes than calamari trawls, only length and age data from finfish (A-, G-, and Wlicences) and experimental (E-licence) vessels were used, to avoid biasing length-frequency distributions if proportionally more samples are recorded from one fishery or another in different years. Skate and IIlex trawls also were excluded because their different targets could relate to characteristically different length-frequency distributions of red cod.

LBI method indicators were then selected and scored using Tables 2.1.1.4.1 and

### 2.1.2.2 in ICES (2015) as templates:

1) Length at half the modal catch length should be bigger than L50, for conservation of immature fish ( $L_{c} / L 50>1$ ). Note that length at half the modal catch length may be poorly defined if the catch length-frequency distribution is not smooth and unimodal.
2) Length at cumulative $25^{\text {th }}$ percentile of catch numbers should be bigger than L50, for conservation of immature fish ( $\mathrm{L}_{25 \%} / \mathrm{L} 50>1$ ).
3) Mean length of the largest $5 \%$ of individuals in the catch should be at least $80 \%$ of the asymptotic length, as a benchmark that enough large individuals are in the stock ( $L_{\text {max5\% }}$ / $\mathrm{L}_{\mathrm{ln} f}>0.8$ ).
4) 'Mega-spawners' should comprise at least $30 \%$ of the catch (thus implicitly represent at least $30 \%$ of the stock), as large, old fish disproportionately benefit the resilience of the population (Froese 2004) ( $\mathrm{P}_{\text {mega }}>0.3$ ). Mega-spawners are defined as individuals larger than optimum length (Lopt) $+10 \%$, where Lopt is described as the length at which growth rate is maximum (ICES 2015), or the length at which total biomass of a year-class reaches its maximum value (Froese \& Binohlan 2000). Lopt $=3 \cdot \mathrm{Lnf}_{\mathrm{nf}} \cdot\left(3+\mathrm{Mk}^{-1}\right)^{-1}$ (Beverton 1992), where M is instantaneous natural mortality, k is the rate of curvature of the von Bertalanffy growth function, and the ratio $\mathrm{Mk}^{-1}$ is set in WKLIFE $V$ software (ICES 2015) at the standard constant of 1.5 (Jensen 1996).
5) Mean length of individuals larger than $L_{c}$ ( $L_{\text {meanc }}$ ) should be approximately equal to Lopt, for optimal yield ( $L_{\text {meanc }} / L_{\text {opt }} \approx 1$ ).
6) $L_{\text {meanc }}$ should be equal or bigger to the length-based proxy for MSY ( $L_{F=M}$ ), for producing maximum sustainable yield ( $L_{\text {meanc }} / L_{F=M} \geq 1$ ). $L_{F=M}$ implements the premise that $M S Y$ is attained when fishing mortality equals natural mortality (Froese et al. 2018), and in WKLIFE V software (ICES 2015) is computed as ( $3 \cdot \mathrm{~L}_{\mathrm{c}}+\mathrm{L}_{\mathrm{Inf}}$ )/4.

Margins of variability of the six indicators were estimated by randomly re-sampling $10,000 \times$ on the normal distribution each year's fits of LInf and L50. Indicators were scored against the 'traffic light' scale (ICES 2015) with reference criteria > 1.0 for conservation of immature fish, $>0.8$ for conservation of large fish, and $>0.3$ for conservation of megaspawners. The score was green if the lower $95 \%$ quantile of the re-sampled iterations was > $1.0,>0.8$, and $>0.3$, yellow if $1.0,0.8$, and 0.3 were between the lower and upper $95 \%$ quantiles, and red if the upper $95 \%$ quantile of the re-sampled iterations was $<1.0,<0.8$, and <0.3. The use of the margins of variability means that same empirical values of indicators may be scored different colours in different years. Reference criterion $\approx 1.0$ for optimal yield was green if the lower and upper 95\% quantiles spanned 1.0, yellow if the lower and upper 95\% quantiles spanned 0.9 (the threshold used in ICES 2015) without spanning 1.0, and red otherwise. Reference criterion $\geq 1.0$ for MSY was scored the same as $>1.0$, except that empirical values $\geq 1.0$ were automatically green.

## Length-age relationship

The von Bertalanffy growth function (R package 'fishmethods'; Nelson 2019) was used to fit all red cod length-at-age data from finfish and experimental vessels. Red cod length and age data were jointly available for years 2002-2019, with status of age data advised 'with caution' (Lee et al. 2020) as verification of these ages is required. Growth model parameters (Llnf, $k$, and $t_{0}$ ) were calculated for females and males using nonlinear least square regression. A likelihood ratio test (R package 'fishmethods'; Nelson 2019) was used to test whether the von Bertalanffy growth function was significantly different between females and males. Variabilities of the growth model parameters were estimated by bootstrapping; residuals of the model fits were randomly re-sampled with replacement, added back to the expected lengths, and re-fit to the von Bertalanffy growth function. The $95 \%$ quantiles of 10,000 iterations were retained as confidence intervals. Inter-annual trend of von Bertalanffy Linf was calculated by LOESS (span $=0.75$, degree $=2$ ).

## Length and age at 50\% maturity

Overall and yearly length at 50\% maturity (L50) was calculated as the mid-point of the binomial logistic regression of maturity ogives vs. length (Heino et al. 2002). Sex and maturity were identified following the fish maturity scale by Brickle et al. (2005; modified from Nikolsky 1963): I) immature; II) resting; III) early developing; IV) late developing; V) ripe; VI) running; VII) spent; VIII) recovering spent. Maturity is cyclical as fish pass from post-spawning phase to resting phase, and definitive maturity assignments can only be made that stages I through III are immature, and stages IV+ are always adult; however, stage II can be uncertain as it can look macroscopically similar to maturity stage VIII (A. Arkhipkin, FIFD, pers. comm.). Therefore, maturity assignment was simplified to a dichotomous classification of 0) juvenile, including maturity stages I and III, and 1) adult, including maturity stages IV to VIII, omitting stage II. Annual L50s were calculated from randomly sampled individuals collected through the FICZ under finfish and experimental licences during the spawning season (AugustOctober; Brickle et al. 2011), from 1992 through 2021. Trends of annual L50 were calculated with LOESS (span $=0.75$, degree $=2$ ). Overall and yearly age at $50 \%$ maturity (A50) was calculated for females and males separately, by predicting age corresponding to L50 using the inversed von Bertalanffy equation.

## Catch at length

Yearly length frequency distributions, from 1990 through 2021, were examined for females and males to describe patterns in catch at length through time. Unsexed individuals were excluded from the analysis. Lengths of individuals sampled randomly and caught by finfish and the experimental vessels through the FICZ from January through December were included in the analysis. Yearly length frequencies were compared with yearly L50 to assess if the catch was mainly comprised of immature or mature individuals.

## Natural mortality

Natural mortality (M) of red cod was calculated as an indicator to examine vulnerability of the stock. Natural mortality is the component of total mortality that is not caused by fishing, but by causes such as predation, diseases, senility, pollution, amongst other factors. Annual natural mortality refers to the proportion of fish dying during the year expressed as a fraction of the fish alive at the beginning of the year (FAO 1999), and was calculated using equation 1 following Then et al. (2015):
$\mathrm{M}=4.899 \times \mathrm{t}_{\text {max }}^{-0.916}$
Eqn. 1
where $t_{\text {max }}=$ maximum age, taken as the oldest age reported in the FIFD database not considered an outlier. Then et al. (2015) recommended the use of the $t_{\max }$-based estimator over other estimators based on cross-validation of prediction error, model residual patterns, model parsimony, and biological considerations.

All analyses were performed in RStudio ( R Core Team 2021).

## Results

## ICES advice rules

## ICES Category 3 Total Allowable Catch

TAC options for 2023 based on the current year TAC were calculated as follows:
TAC_32022 $=1,199$ tonnes.

ICES 2/3 rule weighted by inverse variances:
$T A C_{-} 3_{2022} \times \frac{\left(\frac{\left(35217.39 \times 1.85^{-8}\right)+\left(81176.73 \times 2.27^{-10}\right)}{1.85^{-8}+2.27^{-10}}\right)}{\left(\frac{\left(57422.88 \times 1.21^{-9}\right)+\left(83005.12 \times 2.06^{-9}\right)+\left(21889.98 \times 2.91^{-8}\right)}{1.21^{-9}+2.06^{-9}+2.91^{-8}}\right)}$
$=1582.35$

20\% cap:
$T A C_{32022}+20 \%=1438.8$

Thus:
$T A C_{-} 3_{2023}=$ Minimum $(1582.35,1438.8)=1438.8$ tonnes

## Commercial catch and CPUE

Red cod catches in Falkland Islands waters have averaged 1,671 t per year since 1987, representing approximately $58 \%$ of the Falkland Islands and Argentine combined annual catch (Fig. 1). Falkland Islands and Argentine red cod catches were positively associated (Fig. 2; Appendix I).


Fig. 1. Annual commercial catch of red cod in Falkland Islands, Argentine, and Chilean waters. Falkland Islands commercial catch data exclude experimental (E-licence) and out-of-zone (O-licence) licences from 1990; earlier than 1990 these licences were not designated.


Fig. 2. Falkland Islands vs. Argentina annual commercial catches of red cod from 1987 to 2021, with LOESS smooth $\pm 95 \%$ confidence intervals (LOESS; span $=0.75$, degree $=2$ ).

From 1990 through 2021, approximately $90 \%$ of the annual red cod catch in the FICZ was from finfish licences (A-, G-, and W-licences). During 2021 a total of 1,194 t of red cod were reported caught in Falkland Islands waters, of which $1,187 \mathrm{t}$ were caught under commercial licences, i.e., excluding the experimental E-licence. Approximately $44 \%$ of all Falkland Islands red cod catch was under W-licence, $27 \%$ was under A-licence, and 26\% under G-licence in 2021; the three finfish licences (A-, G-, and W-licences) together accounted for $97 \%$ of the total red cod catch (Table I). Reported red cod discards were $1.5 \%$ of the total red cod catch in 2021; calamari vessels ( C - and X -licences) discarded an average of $19 \%$ of their red cod catch, Illex vessels (B-licence) discarded 100\%, and finfish vessels discarded an average of $0.8 \%$.

Table I. Catch proportion of red cod by licence type in Falkland Islands waters during 2021.

| Licence | Target species | Catch $\text { ( } \mathrm{t} \text { ) }$ | Catch (\%) | Discard (t) | Proportion discarded (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| W | Restricted finfish | 524.579 | 43.93 | 7.702 | 1.47 |
| A | Unrestricted finfish | 321.875 | 26.96 | 2.767 | 0.86 |
| G | Restricted finfish and IIlex | 313.785 | 26.28 | 0.609 | 0.19 |
| C | Calamari $1^{\text {st }}$ season | 17.747 | 1.49 | 2.373 | 13.37 |
| X | Calamari ${ }^{\text {nd }}$ season | 7.131 | 0.60 | 1.757 | 24.64 |
| E | Experimental | 7.015 | 0.59 | 0.931 | 13.27 |
| B | Illex squid | 1.440 | 0.12 | 1.440 | 100.00 |
| $\mathrm{F}^{\text {a }}$ | Skates and rays | 0.000 | 0.00 | 0.000 | 0.00 |
| L | Toothfish (longline) | 0.000 | 0.00 | 0.000 | 0.00 |
| $S^{\text {a }}$ | Southern blue whiting and hoki | 0.000 | 0.00 | 0.000 | 0.00 |
| 0 | Outside Falkland Islands waters | 0.000 | 0.00 | 0.000 | 0.00 |
| Total |  | 1193.572 | 99.97 | 17.579 | 1.47 |

${ }^{\text {a }} \mathrm{F}$ and S licenses were not fished during 2021.

Average CPUE ranged from $53 \mathrm{~kg} / \mathrm{h}$ in 2020 to a maximum of $322 \mathrm{~kg} / \mathrm{h}$ in 1996, although most values were below $250 \mathrm{~kg} / \mathrm{h}$. The CPUE trend showed two modes approximately 15 years apart; CPUE increased from 1990 to reach the highest CPUE in the time series in 1996 ( $322 \mathrm{~kg} / \mathrm{h}$ ), and declined towards 2003. CPUE increased again towards 2013 (232 kg/h) and declined to $55 \mathrm{~kg} / \mathrm{h}$ in 2021 (Fig. 3).


Fig. 3. Yearly CPUE $\pm 1$ standard error of red cod in Falkland Islands waters from 1990 through 2021, calculated from finfish (A-, G-, and W-licences) vessels, with LOESS smooth $\pm 95 \%$ confidence intervals (LOESS; span $=0.75$, degree $=2$ )

The monthly CPUE by finfish licences from 1990 through 2020 ranged between $82 \mathrm{~kg} / \mathrm{h}$ and $140 \mathrm{~kg} / \mathrm{h}$; CPUE increased from January to March, and remained relatively high from March through May. CPUE declined towards July, and increased again from August through October. There was no fishing effort in January and February 2021, CPUE remained relatively similar from March through November and increased significantly in December (Fig. 4; Appendix II). During 2021, red cod were caught mainly to the west of West Falkland, between $50^{\circ} \mathrm{S}$ and $53^{\circ} \mathrm{S}$, and between $62^{\circ} \mathrm{W}$ and $63^{\circ} \mathrm{W}$; minor catches were also reported to the north in the FICZ (Appendix III).


Fig. 4. Monthly average CPUE $\pm 1$ standard error of red cod in Falkland Islands waters from 1990 through 2020 (red), and in 2021 (blue), calculated from finfish (A-, G-, and W-licences) vessels, with LOESS smooths $\pm 95 \%$ confidence intervals (LOESS; span $=0.75$, degree $=2$ ).

## Surveys biomass estimates

## Summer surveys (February)

The biomass of red cod during the February surveys did not change significantly from 2010 to 2022. The biomass of red cod in 2022 ( $81,176.73 \mathrm{t}$ ) was comparable to the biomass in 2010 ( $95,050.09$ t; Table II; Fig. 5). Accordingly, only 5,776 out of 10,000 paired re-samples had higher biomass estimate values in February 2010 than in February 2022 (57.8\%), therefore the difference in biomass between 2022 and 2010 is not significant at $p>0.05$. During the February 2015-2018 surveys, red cod were mainly aggregated to the west edge of the FICZ. The aggregations were less dense and their distribution spread to the north and southwest of West Falkland during the February 2019-2022 surveys, with a decrease in the density of the aggregations compared with 2015-2018 (Appendix IV).

Table II. Summer (February) surveys catch and effort, and biomass estimates (mean $\pm 95 \%$ confidence intervals) of red cod in Falkland Islands waters.

| Year | Survey | Trawls (n) | Swept area ( $\mathrm{km}^{2}$ ) | Effort <br> (h) | Catch (kg) | $\begin{aligned} & \text { CPUE } \\ & (\mathrm{kg} / \mathrm{h}) \end{aligned}$ | Biomass <br> ( t ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | Groundfish | 87 | 17.04 | 87.52 | 13427.48 | 153.43 | 95050.09 |
|  | D. gahi | 55 | 42.29 | 109.27 | 111.60 | 1.02 | (18335.99-158897.80) |
|  | Total | 142 | 59.34 | 196.78 | 13539.08 | 68.80 |  |
| 2011 | Groundfish | 88 | 17.21 | 88.00 | 23099.27 | 262.49 | $\begin{gathered} 166617.50 \\ (39230.31-258711.16) \end{gathered}$ |
|  | D. gahi | 58 | 40.04 | 110.63 | 440.27 | 3.98 |  |
|  | Total | 146 | 57.26 | 198.63 | 23539.54 | 118.51 |  |
| 2015 | Groundfish | 89 | 16.72 | 90.17 | 20314.03 | 225.29 | $\begin{gathered} 106244.23 \\ (45278.81-160780.36) \end{gathered}$ |
|  | D. gahi | 57 | 46.90 | 111.50 | 1495.40 | 13.41 |  |
|  | Total | 146 | 63.61 | 201.67 | 21809.43 | 108.15 |  |
| 2016 | Groundfish | 90 | 17.64 | 91.42 | 18644.48 | 203.95 | $\begin{gathered} 102789.02 \\ (28384.22-149860.74) \end{gathered}$ |
|  | D. gahi | 56 | 54.46 | 107.92 | 1302.61 | 12.07 |  |
|  | Total | 146 | 72.10 | 199.33 | 19947.08 | 100.07 |  |
| 2017 | Groundfish | 90 | 18.52 | 92.00 | 11104.46 | 120.70 | $\begin{gathered} 59568.95 \\ (22863.35-86532.41) \end{gathered}$ |
|  | D. gahi | 58 | 54.09 | 117.00 | 2717.14 | 23.22 |  |
|  | Total | 148 | 72.62 | 209.00 | 13821.59 | 66.13 |  |
| 2018 | Groundfish ${ }^{\text {a }}$ | 97 | 20.47 | 96.42 | 12733.50 | 132.07 | $\begin{gathered} 57422.88 \\ (19277.51-117355.42) \end{gathered}$ |
|  | D. gahi | 59 | 36.87 | 100.83 | 567.39 | 5.63 |  |
|  | Total | 156 | 57.35 | 197.25 | 13300.89 | 67.43 |  |
| 2019 | Groundfish | 79 | 17.22 | 79.00 | 10652.83 | 134.85 | $\begin{gathered} 83005.12 \\ (35235.62-119480.37) \end{gathered}$ |
|  | D. gahi | 52 | 72.70 | 97.05 | 3029.02 | 31.21 |  |
|  | Total | 131 | 89.93 | 176.05 | 13681.85 | 77.72 |  |
| 2020 | Groundfish ${ }^{\text {a }}$ | 80 | 17.04 | 79.95 | 3334.18 | 41.70 | $\begin{gathered} 21889.98 \\ (10993.21-32014.04) \end{gathered}$ |
|  | D. gahi | 59 | 86.80 | 112.52 | 373.27 | 3.32 |  |
|  | Total | 139 | 103.84 | 192.47 | 3707.45 | 19.26 |  |
| 2021 | Groundfish | 80 | 16.43 | 79.48 | 5681.47 | 71.48 | $\begin{gathered} 35217.39 \\ (22852.74-51663.11) \end{gathered}$ |
|  | D. gahi | 55 | 90.65 | 111.22 | 358.65 | 3.22 |  |
|  | Total | 135 | 107.07 | 190.70 | 6040.12 | 31.67 |  |
| 2022 | Groundfish | 42 | 9.22 | 41.90 | 9595.83 | 229.02 | $\begin{gathered} 81176.73 \\ (34162.13-129660.26) \end{gathered}$ |
|  | D. gahi | 60 | 86.75 | 119.08 | 817.52 | 6.87 |  |
|  | Total | 102 | 95.97 | 160.98 | 10413.34 | 64.69 |  |

${ }^{2}$ An additional one-day transect of four trawls was taken in shallow inshore waters to sample for juvenile toothfish. These four trawls were not included in the analyses as their locations were not relevant to the distribution of red cod. Groundfish February surveys were not conducted in 2012, 2013, and 2014.


Fig. 5. Red cod biomass estimates (red dots) $\pm 95 \%$ confidence intervals from February surveys in Falkland Islands waters, with LOESS smooths $\pm 95 \%$ confidence intervals (LOESS; span $=1$, degree $=2$ ). No parallel February surveys (groundfish and Patagonian squid pre-season surveys) were conducted in 2012, 2013, and 2014.

## Winter surveys (July)

The estimated biomass of red cod in the July 2020 survey ( $38,777 \mathrm{t}$ ) was $3 \times$ greater than in the July 2017 surveys ( 12,480 ; Table III). However, only 7,349 out of 10,000 paired resamples had higher biomass estimate values in July 2020 than in July 2017 (73.5\%), thus not significant at $p>0.05$. In July 2017, aggregations of red cod were detected to the north and to the west in the FICZ, whereas in July 2020 red cod were mainly aggregated to the northwest limit of the FICZ (Appendix V). Differences in biomass estimates between February and July surveys are likely due to the migratory pattern of red cod.

Table III. Winter (July) surveys catch and effort, and biomass estimates (mean $\pm 95 \%$ confidence intervals) of red cod in Falkland Islands waters.

| Year | Survey | Trawls (n) | Swept area ( $\mathrm{km}^{2}$ ) | Effort <br> (h) | Catch (kg) | $\begin{aligned} & \text { CPUE } \\ & (\mathrm{kg} / \mathrm{h}) \end{aligned}$ | Biomass <br> ( t ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | Groundfish | 74 | 15.40 | 74 | 1784.05 | 24.11 | $\begin{gathered} 12480.10 \\ (6345.05-18207.66) \end{gathered}$ |
|  | D. gahi $^{\text {a }}$ | 59 | 54.70 | 114 | 779.17 | 6.83 |  |
|  | Total | 133 | 70.10 | 188 | 2563.22 | 13.63 |  |
| 2020 | Groundfish ${ }^{\text {b }}$ | 33 | 7.14 | 33 | 2164.71 | 65.59 | $\begin{gathered} 38776.73 \\ (6143.58-68995.25) \end{gathered}$ |
|  | D. gahi | 55 | 98.57 | 101 | 419.05 | 4.15 |  |
|  | Total | 88 | 105.71 | 134 | 2583.76 | 19.28 |  |

${ }^{\text {a }}$ An additional one-day transect of four trawls was taken in shallow inshore waters to sample for juvenile toothfish. These four trawls were not included in analyses as their locations were not relevant to the distribution of red cod.
${ }^{\mathrm{b}}$ Twelve additional trawls were conducted in high seas during the July 2020 survey; these trawls were not included in the analyses.
Note that no parallel July surveys (groundfish and Patagonian squid pre-season surveys) were conducted in 2018 and 2019.

## Length and age analyses

## Length Based Indicators

Yearly 'traffic light' length indicators for females and males are summarized in Table IV. Indicator Lc/L50, for conservation of immature fish, had positive outcomes (green) most years from 1992 to 2004, and had negative outcomes (red) most years from 2005 to 2021 for females. For males, this indicator was positive in 1996-1997, 2001-2002, and in 2008 but it was negative most years since 2003. Indicator $L_{25 \%} / L 50$, also for conservation of immature fish, had variable outcomes from 1992 to 2004 for females and males; however, negative outcomes were common since 2005. Indicator $L_{\max 5 \%} / L_{\operatorname{lnf}}$, for the conservation of large individuals, was of concern (yellow) from 2002 to 2009, 2013, and 2018-2019, and it was negative most years since 2010 for females. For males, this indicator was of concern most years from 2002 to 2021, except for 2005, 2012, 2014, and 2016 with negative outcomes. Indicator $P_{\text {mega, }}$ for the presence of mega-spawners, was of concern most years from 2002 to 2008 for females, and in 2002 and 2008 for males; negative outcomes were common most years since 2009 for females, and since 2003 for males.

Indicator $L_{\text {meanc }} / L_{\text {opt }}$, for optimal yield, had positive outcomes most years from 2002 to 2008 for females, and fluctuated between positive and of concern for males from 2002 to 2008; negative outcomes were common most years for both females and males since 2009. Indicator $L_{\text {meanc }} / L_{F=M}$, for maximum sustainable yield, was mostly positive from 2002 to 2021
for both females and males, except for a few years after 2009. Some LBI indicator outputs (Table IV) differ from the 2021 red cod stock assessment (Ramos and Winter 2021) due to length-age measurements from bottom trawl experimental (E-licence) being included in the current assessment, but not measurements from semi-pelagic trawls as in 2021.

Table IV. Red cod indicators by sex and year, with 'traffic light' scoring. Lc) Length at half the modal catch length; L50) Length at $50 \%$ maturity; $\mathrm{L}_{25 \%}$ ) Length at cumulative $25^{\text {th }}$ percentile of catch; $\mathrm{L}_{\text {max5\% }}$ ) Mean length of the largest $5 \%$ of individuals in the catch; Llnff Asymptotic average maximum body size; $P_{\text {mega) }}$ ) Proportion of 'Mega-spawners' in the catch; $\mathrm{L}_{\text {meanc }}$ ) Mean length of individuals larger than LC; $L_{\text {opt }}$ ) Optimum length; $\mathrm{L}_{\mathrm{F}}=\mathrm{m}$ ) Length-based proxy for MSY. Data were not available in some years (blank cells).


Table IV. continued...

| Sex | Year | Conservation |  |  |  | $\begin{aligned} & \hline \text { Optimal } \\ & \text { yield } \\ & \text { L meanc }^{2} / \text { Lopt }^{\text {opt }} \\ & \approx 1 \end{aligned}$ | $\begin{gathered} M S Y \\ L_{\text {meanc }} / L_{F=M} \\ \geq 1 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{gathered} \mathrm{L}_{\mathrm{c}} / \mathrm{L} 50 \\ >1 \end{gathered}$ | $\begin{gathered} \mathrm{L}_{25 \%} / \mathrm{L} 50 \\ >1 \end{gathered}$ | $\begin{aligned} & \mathrm{L}_{\operatorname{max5} 5} / \mathrm{L}_{\operatorname{lnf}} \\ & \quad>0.8 \end{aligned}$ | $\begin{aligned} & P_{\text {mega }} \\ & >0.3 \end{aligned}$ |  |  |
|  | 1992 | 0.88 | 0.88 |  |  |  |  |
|  | 1993 | 0.97 | 0.93 |  |  |  |  |
|  | 1994 |  |  |  |  |  |  |
|  | 1995 |  |  |  |  |  |  |
|  | 1996 | 1.14 | 1.16 |  |  |  |  |
|  | 1997 | 1.15 | 1.11 |  |  |  |  |
|  | 1998 | 0.96 | 0.98 |  |  |  |  |
|  | 1999 |  |  |  |  |  |  |
|  | 2000 | 0.89 | 0.89 |  |  |  |  |
|  | 2001 | 1.24 | 1.31 |  |  |  |  |
|  | 2002 | 1.27 | 1.20 | 0.91 | 0.36 | 1.12 | 1.05 |
|  | 2003 | 0.75 | 0.95 | 0.77 | 0.05 | 0.80 | 1.03 |
|  | 2004 | 0.72 | 0.91 | 0.87 | 0.19 | 0.85 | 1.13 |
|  | 2005 | 0.56 | 0.86 | 0.71 | 0.01 | 0.71 | 1.05 |
| M | 2006 | 1.01 | 1.03 | 0.84 | 0.09 | 0.91 | 0.97 |
|  | 2007 | 0.51 | 0.58 | 0.79 | 0.08 | 0.80 | 1.21 |
|  | 2008 | 1.10 | 1.06 | 0.84 | 0.22 | 1.05 | 1.03 |
|  | 2009 | 0.44 | 0.55 | 0.86 | 0.17 | 0.76 | 1.19 |
|  | 2010 | 0.85 | 0.87 | 0.75 | 0.02 | 0.82 | 0.93 |
|  | 2011 | 0.50 | 0.68 | 0.75 | 0.02 | 0.72 | 1.08 |
|  | 2012 | 0.46 | 0.58 | 0.63 | 0.00 | 0.55 | 0.86 |
|  | 2013 | 0.90 | 0.93 | 0.81 | 0.12 | 0.90 | 1.03 |
|  | 2014 | 0.28 | 0.30 | 0.65 | 0.01 | 0.45 | 0.86 |
|  | 2015 | 0.51 | 0.76 | 0.81 | 0.04 | 0.73 | 1.11 |
|  | 2016 | 0.86 | 0.84 | 0.69 | 0.00 | 0.81 | 0.93 |
|  | 2017 |  |  |  |  |  |  |
|  | 2018 | 0.69 | 0.78 | 0.67 | 0.01 | 0.74 | 0.94 |
|  | 2019 | 0.46 | 0.71 | 0.61 | 0.01 | 0.64 | 0.97 |
|  | 2020 | 0.78 | 0.72 |  |  |  |  |
|  | 2021 | 0.44 | 0.54 |  |  |  |  |

## Length-age relationship

The length-age relationship of females and males pooled ( $n=8,283$ ) gave the values: $L_{\text {Inf }}=100.77 \mathrm{~cm}, k=0.0934$, and $t_{0}=-0.5937$ years. Length and age of females ( $n=4,930$ ) ranged from 9 cm to 90 cm , and from 1 year to 24 years, respectively. The length-age relationship of females gave the values: $\mathrm{L}_{\mathrm{Inf}}=103.79 \mathrm{~cm}, \mathrm{k}=0.0918$, and $\mathrm{t}_{0}=-0.5363$ years. Length and age of males $(n=3,353)$ ranged from 9 cm to 86 cm and from 1 year to 22 years, respectively. The length-age relationship of males gave the values: $L_{\operatorname{lnf}}=89.95 \mathrm{~cm}, \mathrm{k}=0.1070$,
and $t_{0}=-0.5984$ years (Appendix VI). Yearly von Bertalanffy parameters are summarized in Appendix VII. Asymptotic lengths (LInf) did not change significantly from 2002 to 2019 (Fig. 6).


Fig. 6. Asymptotic lengths $\left(L_{\operatorname{lnf}}\right) \pm 1$ standard error calculated according to the von Bertalanffy growth function for female (red dots) and male (blue dots) red cod caught by finfish (A-, G-, and W-licences) and experimental (E-licence) vessels in the FICZ through the year, from 2002 through 2019, with LOESS smooths $\pm 95 \%$ confidence intervals (LOESS; span $=0.75$, degree $=2$ ).

## Length and age at 50\% maturity

Over the time series 1992-2021, length at 50\% maturity (L50) of females was $50.9 \pm$ 0.24 cm total length ( $n=10,594$ ) and age at $50 \%$ maturity (A50) was 6.8 years old; L50 of males was $44.5 \pm 0.18 \mathrm{~cm}$ total length ( $n=13,083$ ) and A50 was 5.8 years old. Therefore, immature females are inferred as $<7$ years old and mature females are inferred as $\geq 7$ years
old; immature males are inferred as $<6$ years old and mature males are inferred as $\geq 6$ years old. Annual L50 and A50 of females ranged from 33.6 cm and 3.7 years old in 2006 to 62.4 cm and 9.5 years old in 2011, respectively. Annual L50 and A50 of males ranged from 32.6 cm and 3.6 years old in 2006 to 68.7 cm and 12.9 years old in 2020. The L50 fit did not change significantly for females from 1992 through 2021 but it increased significantly for males in recent years (Fig. 7; Appendixes VIII-IX).


Fig. 7. Lengths at $50 \%$ maturity (L50) $\pm 1$ standard error of female (red dots) and male (blue dots) red cod caught by finfish (A-, G-, and W-licences) and experimental (E-licence) vessels in the FICZ from August through October, from 1992 through 2021, with LOESS smooths $\pm 95 \%$ confidence intervals (LOESS; span $=0.75$, degree $=2$ ).

## Catch at length

Female red cod ( $\mathrm{n}=85,555$; 1990 to 2021) ranged from 9 cm to 97 cm total length, and males ( $\mathrm{n}=60,172$; 1990 to 2021) ranged from 9 cm to 87 cm total length. Length-groups were not discernible due to size overlap most years. Females and males were mostly caught at $>40 \mathrm{~cm}$ total length from 2002 through 2008, smaller length-groups (< 40 cm total length) became dominant from about 2009 through 2021 (Fig. 8). The catch was mostly comprised of females and males at sizes $\leq$ L50 in $63 \%$ and $73 \%$ of the total number of years assessed ( $\mathrm{n}=$ 32 ), respectively. Catch at length was $\geq$ L50 during a few years before 2008 , and it was $\leq L 50$ from 2009 onwards (Fig. 8; Appendix X).


Fig. 8. Length frequency distribution of female and male red cod caught by finfish (A-, G-, and Wlicences) and experimental (E-licence) vessels in the FICZ from 1990 through 2021. The black solid lines indicate the length at 50\% maturity (L50). Note the gap in 1994 due to limited data.

## Natural mortality

Equation 1 resulted in a natural mortality ( M ) calculation of:
$\mathrm{M}=4.899 \times \mathrm{t}_{\max }^{-0.916}=4.899 \times 24^{-0.916}=0.2666$
indicating that $26.6 \%$ of the stock dies per year not by fishing but due to natural causes such as predation, diseases, senility, amongst others.

## Conclusions

Red cod Total Allowable Catch for 2023 was set at $1,439 \mathrm{t}$, calculated using the ICES category 3 framework, which represents an increase of $18 \%$ of the total commercial red cod catch in 2021 ( $1,186.6 \mathrm{t}$ ), and an increase of 20\% of the TAC for 2022 ( $1,199 \mathrm{t}$ ).

Most of the red cod catch (97\%) in Falkland Islands waters in 2021 was taken between the three finfish licences (A-, G-, and W-licences), with the greatest contribution to the catch by W-licence (44\%).

Red cod commercial CPUE in the FICZ varied from 1990 to 2021; CPUE increased from 1990 through 1996, followed by a significant decline from 1997 through 2003. CPUE increased again towards 2013 and declined towards 2021. Intra-annually, the highest CPUE of red cod occurred from March through May, and from August through October.

February surveys biomasses showed no significant change in red cod abundance from 2010 through 2022; however, February surveys likely reflect variability in the migratory timing of this species. The 2017 and 2020 July surveys revealed comparable red cod biomasses, with overlapping confidence intervals.

Length-based indicators suggest that conservation of immature fish, large individuals, mega-spawners, and optimal yield were of concern or negative most years, whereas MSY was of concern or negative since about 2010. However, red cod spawning grounds to the south and southwest of West Falkland were closed for fishing since 2010 and large individuals were not targeted (A. Arkhipkin pers. comm.). This may have resulted in the decrease of CPUE during the months when spawning aggregations occur, i.e., August through October, which may also affect the LBI outcomes for large individuals and for mega-spawners encountered over the past decade.

The length and age analyses showed no significant change for Linf from 2002 to 2019, and of length at $50 \%$ maturity of females from 1992 to 2021. Comparison of length at $50 \%$ maturity and catch at length revealed that red cod were mostly caught before reaching
maturity, which can reduce stock sustainability (Vasilakopoulos et al. 2011; Muluye et al. 2016; Ben-Hasan et al. 2021).

The multiple analyses used in this study suggest that the red cod stock is currently in poor condition and conservation measures should be implemented. Control of fishing pressure, and of by-catch and discard of small individuals of no commercial value should be of high importance in Falkland Islands fisheries given the trends detected.

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## Appendix

Appendix I. Annual commercial catches ( t ) of red cod reported in Falkland Islands (excluding E-licence; Falkland Islands Governmente; Falkland Islands Government 2021) and Argentina (Argentine Government ${ }^{f}$; Sánchez et al. 2012; Navarro et al. 2014, 2019).

| Year | Falkland Islands (t) | Argentina (t) |
| ---: | ---: | ---: |
| 1987 | 88.0 | 46.9 |
| 1988 | 5121.0 | 47.5 |
| 1989 | 2817.0 | 1186.0 |
| 1990 | 2778.3 | 2115.5 |
| 1991 | 2879.6 | 1272.2 |
| 1992 | 7056.9 | 3050.0 |
| 1993 | 6230.6 | 2207.3 |
| 1994 | 4042.8 | 1310.4 |
| 1995 | 9085.0 | 2359.3 |
| 1996 | 6960.5 | 2077.0 |
| 1997 | 4691.3 | 2610.4 |
| 1998 | 8028.4 | 6808.5 |
| 1999 | 9234.9 | 7202.9 |
| 2000 | 6556.3 | 9431.3 |
| 2001 | 3896.2 | 4449.0 |
| 2002 | 2617.2 | 3129.1 |
| 2003 | 2283.9 | 5689.1 |
| 2004 | 2779.6 | 4664.3 |
| 2005 | 2465.1 | 3185.5 |
| 2006 | 3440.2 | 2962.0 |
| 2007 | 5191.6 | 4609.8 |
| 2008 | 4075.3 | 8009.5 |
| 2009 | 5093.7 | 6962.7 |
| 2010 | 3098.9 | 6813.0 |
| 2011 | 4184.2 | 5190.7 |
| 2012 | 4590.5 | 3921.9 |
| 2013 | 5103.4 | 3814.9 |
| 2014 | 3447.0 | 2780.1 |
| 2015 | 3312.4 | 2289.2 |
| 2016 | 3122.1 | 2008.3 |
| 2017 | 1363.0 | 1511.2 |
| 2018 | 1637.5 | 977.6 |
| 2019 | 1725.5 | 396.6 |
| 2020 | 1413.7 | 685.8 |
| 2021 | 1186.6 | 714.5 |
|  |  |  |
|  |  |  |

[^2]Appendix II. Monthly CPUE of red cod in Falkland Islands waters from 1990 to 2021, calculated from finfish (A-, G-, and W-licences) vessels, with LOESS smooths $\pm 95 \%$ confidence intervals (LOESS; span $=0.75$, degree $=2$ ).


Appendix III. Monthly CPUE of red cod in Falkland Islands waters during 2021, calculated from finfish (A-, G-, and W-licences) vessels. There was no fishing effort during January and February under finfish licences.






## Appendix III. continued...



Appendix IV. Densities of red cod modelled by inverse distance weighting in the FICZ, during the February 2010-2022 groundfish and Patagonian squid pre-season surveys.


Appendix IV. continued...


Appendix V. Densities of red cod modelled by inverse distance weighting in the FICZ, during the July 2017 and July 2020 groundfish and Patagonian squid pre-season surveys.



Appendix VI. von Bertalanffy age-length relationship of female and male red cod collected in finfish ( $\mathrm{A}-\mathrm{G}$ - , and W -licences) and experimental ( E -licence) vessels in the FICZ. Age was determined by MFRI ( $n=8,207$ ) and FIFD $(n=76)$ staff.



Appendix VII. Red cod von Bertalanffy length-at-age parameters for curvature (k), age of fish at length zero ( $\mathrm{t}_{0}$ ), and asymptotic length ( $\mathrm{L}_{\mathrm{nf}}$ ), by year and sex, with $95 \%$ confidence intervals. Red cod were collected in finfish (A-, G-, and W-licences) and experimental (E-licence) vessels in the FICZ.

|  | Year | N |  | K | $\mathrm{t}_{0}$ (years) |  | $\mathrm{L}_{\text {Inf }}(\mathrm{cm})$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2002 | 76 | 0.091 | (0.042-0.143) | -0.749 | (-2.831-0.495) | 102.6 | ( 86.3-154.1) |
|  | 2003 | 123 | 0.110 | (0.071-0.154) | -0.583 | (-1.400-0.052) | 92.3 | ( 80.3-115.7) |
|  | 2004 | 281 | 0.085 | (0.065-0.104) | -0.892 | (-1.473-0.427) | 108.9 | ( 99.3-124.6) |
|  | 2005 | 241 | 0.105 | (0.088-0.121) | -0.262 | (-0.619-0.046) | 99.1 | ( 92.9-107.5) |
|  | 2006 | 172 | 0.123 | (0.100-0.147) | -0.505 | (-1.012-0.089) | 90.4 | ( 84.5-98.5) |
|  | 2007 | 74 | 0.113 | (0.084-0.144) | -0.370 | (-1.112-0.224) | 89.7 | ( 81.8-101.6) |
|  | 2008 | 334 | 0.092 | (0.076-0.108) | -0.768 | (-1.141-0.445) | 100.6 | ( 93.2-110.9) |
|  | 2009 | 336 | 0.077 | (0.062-0.093) | -0.710 | (-1.060-0.409) | 116.0 | (105.2-131.9) |
|  | 2010 | 204 | 0.075 | (0.048-0.102) | -0.984 | (-1.603-0.511) | 117.6 | ( 99.8-155.5) |
| F | 2011 | 371 | 0.065 | (0.050-0.080) | -0.869 | (-1.218-0.568) | 130.1 | (114.7-153.8) |
|  | 2012 | 196 | 0.086 | (0.071-0.101) | -0.618 | (-0.988-0.296) | 105.7 | ( 97.7-116.8) |
|  | 2013 | 88 | 0.088 | (0.050-0.126) | -0.865 | (-1.663-0.302) | 96.0 | ( 79.9-136.8) |
|  | 2014 | 457 | 0.066 | (0.053-0.081) | -0.719 | $(-0.935-0.524)$ | 128.1 | (113.1-150.3) |
|  | 2015 | 423 | 0.094 | (0.081-0.106) | -0.589 | (-0.792--0.41) | 102.2 | ( 95.6-111.0) |
|  | 2016 | 373 | 0.102 | (0.085-0.119) | -0.222 | (-0.465-0.007) | 98.1 | ( 90.5-108.3) |
|  | 2017 | 222 | 0.090 | (0.074-0.106) | -0.283 | (-0.527-0.067) | 107.4 | ( 97.9-120.5) |
|  | 2018 | 167 | 0.053 | (0.025-0.081) | -0.946 | (-1.714--0.35) | 144.8 | (111.4-246.9) |
|  | 2019 | 288 | 0.096 | (0.074-0.119) | -0.724 | (-1.156-0.354) | 96.3 | ( 87.1-110.3) |


| Year | N | K | $\mathrm{t}_{0}($ years $)$ |  | $\mathrm{L}_{\operatorname{lnf}}(\mathrm{cm})$ |  |  |  |
| ---: | ---: | ---: | :---: | ---: | :--- | ---: | ---: | ---: |
| 2002 | 62 | 0.081 | $(0.038-0.123)$ | -1.085 | $(-2.730--0.075)$ | 106.2 | $(88.3-166.6)$ |  |
| 2003 | 91 | 0.112 | $(0.074-0.159)$ | -0.977 | $(-1.863--0.206)$ | 84.4 | $(72.8-103.6)$ |  |
| 2004 | 193 | 0.105 | $(0.083-0.126)$ | -0.908 | $(-1.526--0.397)$ | 90.0 | $(83.7-99.3)$ |  |
| 2005 | 159 | 0.065 | $(0.034-0.099)$ | -1.486 | $(-2.392--0.796)$ | 118.4 | $(92.8-181.4)$ |  |
| 2006 | 130 | 0.104 | $(0.075-0.134)$ | -1.345 | $(-2.171--0.690)$ | 87.3 | $(78.5-101.3)$ |  |
| 2007 | 48 | 0.101 | $(0.063-0.138)$ | -0.838 | $(-1.778--0.195)$ | 90.2 | $(78.9-115.6)$ |  |
| 2008 | 195 | 0.090 | $(0.072-0.111)$ | -1.122 | $(-1.560--0.727)$ | 96.2 | $(87.6-108.3)$ |  |
| 2009 | 226 | 0.086 | $(0.067-0.106)$ | -1.059 | $(-1.504--0.683)$ | 98.9 | $(89.6-112.9)$ |  |
| 2010 | 123 | 0.160 | $(0.127-0.191)$ | -0.210 | $(-0.589-0.076)$ | 74.9 | $(69.8-82.8)$ |  |
| M | 2011 | 212 | 0.127 | $(0.107-0.147)$ | -0.346 | $(-0.599--0.125)$ | 82.9 | $(77.4-90.3)$ |
| 2012 | 168 | 0.109 | $(0.093-0.125)$ | -0.370 | $(-0.669--0.110)$ | 91.9 | $(86.5-98.9)$ |  |
| 2013 | 46 | 0.144 | $(0.083-0.208)$ | -0.311 | $(-1.057-0.208)$ | 74.0 | $(62.7-100.1)$ |  |
| 2014 | 350 | 0.077 | $(0.058-0.097)$ | -0.521 | $(-0.713--0.355)$ | 117.7 | $(101.0-145.5)$ |  |
| 2015 | 308 | 0.106 | $(0.088-0.124)$ | -0.568 | $(-0.759--0.395)$ | 91.9 | $(84.0-102.8)$ |  |
| 2016 | 237 | 0.129 | $(0.111-0.148)$ | -0.319 | $(-0.571--0.093)$ | 80.6 | $(75.7-86.7)$ |  |
| 2017 | 157 | 0.090 | $(0.062-0.119)$ | -0.683 | $(-1.094--0.346)$ | 94.9 | $(81.6-119.3)$ |  |
| 2018 | 131 | 0.073 | $(0.036-0.111)$ | -0.971 | $(-1.849--0.343)$ | 108.6 | $(86.3-176.2)$ |  |
| 2019 | 224 | 0.082 | $(0.054-0.112)$ | -1.134 | $(-1.810--0.611)$ | 100.9 | $(86.6-128.3)$ |  |

Appendix VIII. Binomial logistic regressions of juvenile (0) or adult (1) maturity ogives vs. length for female red cod sampled randomly in finfish ( $\mathrm{A}-\mathrm{G}$ - , and W -licences) and experimental (E-licence) vessels in the FICZ. Red lines indicate the intercept for length at 50\% adulthood, corresponding to Fig. 7.


Appendix IX. Binomial logistic regressions of juvenile (0) or adult (1) maturity ogive vs. length for male red cod sampled randomly in finfish ( $\mathrm{A}-, \mathrm{G}$-, and W -licences) and experimental ( E -licence) vessels in the FICZ. Red lines indicate the intercept for length at $50 \%$ adulthood, corresponding to Fig. 7.


Appendix X. Number of red cod individuals sampled for length frequency distributions, corresponding to individuals caught randomly by finfish ( $\mathrm{A}-\mathrm{G}, \mathrm{G}$, and W -licences) and experimental (E-licence) vessels through the year in the FICZ from 1990 to 2021.

| Year | Females (n) | Males (n) |
| :---: | :---: | :---: |
| 1990 | 821 | 679 |
| 1991 | 341 | 309 |
| 1992 | 1322 | 961 |
| 1993 | 469 | 489 |
| 1994 | 0 | 0 |
| 1995 | 11 | 21 |
| 1996 | 843 | 972 |
| 1997 | 2272 | 1527 |
| 1998 | 824 | 606 |
| 1999 | 900 | 884 |
| 2000 | 1297 | 1034 |
| 2001 | 701 | 345 |
| 2002 | 1819 | 1288 |
| 2003 | 814 | 551 |
| 2004 | 3500 | 2459 |
| 2005 | 1566 | 1243 |
| 2006 | 2078 | 1666 |
| 2007 | 2374 | 1506 |
| 2008 | 1704 | 1075 |
| 2009 | 5882 | 4268 |
| 2010 | 3775 | 2360 |
| 2011 | 7050 | 4445 |
| 2012 | 3793 | 2490 |
| 2013 | 4046 | 2594 |
| 2014 | 3163 | 2606 |
| 2015 | 4265 | 3101 |
| 2016 | 5723 | 4511 |
| 2017 | 5021 | 3951 |
| 2018 | 4033 | 2523 |
| 2019 | 4201 | 2795 |
| 2020 | 5668 | 3732 |
| 2021 | 5279 | 3181 |
|  |  |  |


[^0]:    ${ }^{\text {a }}$ http://www.fig.gov.fk/fisheries/publications/fishery-statistics
    ${ }^{\mathrm{b}}$ https://www.agroindustria.gob.ar/sitio/areas/pesca maritima/desembarques/

[^1]:    ${ }^{\text {a http://www.fig.gov.fk/fisheries/publications/fishery-statistics }}$
    ${ }^{\mathrm{b}}$ https://www.agroindustria.gob.ar/sitio/areas/pesca maritima/desembarques/

[^2]:    ${ }^{\mathrm{e}}$ http://www.fig.gov.fk/fisheries/publications/fishery-statistics
    ${ }^{\dagger}$ https://www.agroindustria.gob.ar/sitio/areas/pesca maritima/desembarques/

